

Joint field experiments for comparisons of measuring methods of photosynthetic production*

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Abstract. During the 1st GAP Workshop at Konstanz in April 1982 comparative measurements of phytoplankton primary production by several techniques were conducted simultaneously at an off-shore station in Lake Konstanz and an experimental algal pond. Suspended glass bottle exposure techniques using ¹⁴C and ¹³C uptake gave P_z ($\text{mg C m}^{-3} \text{ h}^{-1}$) values which varied considerably near-surface, but estimates of areal rates for the euphotic zone ΣP_{eu} ($\text{mg C m}^{-2} \text{ h}^{-1}$) which were reasonably close. In the lake, ΣP_z from a vertical tube exposure (with ¹⁴C uptake) was greater than rates derived for integrated bottle samples. The oxygen bottle method permitted a good estimate of compensation depth, corresponding to *in situ* growth studies. There were difficulties in direct comparison between O₂ and carbon methods. Correlation between them for P_z was good in the lake but poor in the pond, both for suspended bottle and vertical tube methods. This series demonstrates that despite reasonable overall estimates, comparatively minor methodological differences in experimental technique can cause large variation.

Introduction

Phytoplankton photosynthesis is a fundamental process in aquatic environments and quantitative information on primary production in a given water body is essential to understanding ecosystem metabolism and the dynamics of phytoplankton populations. The main techniques for quantifying primary production have remained the light and dark oxygen bottle method (Gaarder and Gran, 1927) and the ¹⁴C method (Steemann Nielsen, 1952). Each of these has advantages for certain aquatic environments, but the interpretation of results and an evaluation of their physiological and ecological importance is often difficult. This situation frequently hampers comparison and further analysis of data derived by different

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methods and researchers (Strickland, 1960; Vollenweider, 1974; Fogg, 1975; Harris, 1980; Peterson, 1980).

Recent advances in understanding the biophysics and biochemistry of photosynthesis have also led to further examination of implications of data obtained by routine methods in natural environments (Harris and Piccinin, 1977; Sheldon and Sutcliffe, 1978; Peterson, 1980; Platt, 1980).

The purpose of the joint field experiments, which were conducted during the GAP Workshop at Konstanz, was to compare several methods for measuring phytoplankton photosynthesis and to assess their reliability and limitations in primary production studies. Most participants in these experiments used their own apparatus and techniques.

Experimental locations and materials

The experiments were conducted at two locations on 21 April 1982; a station about 500 m off the northeastern coast of Lake Konstanz, and an experimental pond at the Limnological Institute, University of Konstanz. The depth of the lake station was 80 m and the Secchi disk reading was 4.5 m. Due to cold and stormy weather which prevailed for some days before the start of experiments, the lake water was well mixed vertically with only a slight decrease of water temperature with depth in the euphotic zone (Table I). The dominant phytoplankton were *Rhodomonas lens*, *R. minuta*, *Asterionella formosa*, *Stephanodiscus hantzschii* and μ -algae (Heaney and Sommer, 1984). Chlorophyll *a* concentration, determined spectrophotometrically in hot-ethanol (Marker *et al.*, 1980) and corrected for phaeophytin (Lorenzen, 1967), and dissolved inorganic carbon (DIC) measured by alkalinity titration (Gächter *et al.*, 1984) were almost uniform throughout the euphotic zone (Table I). The 1% depth of the surface irradiance, I_0 , assumed to correspond roughly to the daily compensation depth, was 12.5 m (Jewson *et al.*, 1984).

In the experimental pond, phytoplankton biomass was 14.4 mm³ cell volume l⁻¹ and chlorophyll *a* concentration was 12.5 mg Chl *a* m⁻³ (corrected for phaeopigments) about seven times higher than that in the lake. The phytoplankton was dominated by *Chlamydomonas*, which accounted for 94.3% of total algal biomass. Photosynthetically available radiation (PAR) at the bottom of the pond (depth 1.1 m) was 51% of I_0 and DIC concentration was 9.86 g C m⁻³.

The lake samples were collected with a 9 l PVC Van-Dorn sampler on board the RV Robert Lauterborn, from 0, 1, 2, 3, 5, 8, 10 and 15 m depths. Each sample was transferred into a large plastic bottle (20 l), and after thorough mixing, dispensed into experimental bottles in diffused dim light in the ship's cabin. At the pond, an integrated water column sample down to 1 m depth was taken with a plexiglass tube, mixed well in a carboy and distributed into experimental bottles in the laboratory.

Methods

Types of exposure experiments

Two types of experimental enclosures, glass bottles and vertical perspex tubes,

Table I. Lake Konstanz: physico-chemical and biological background data at the experimental station on 21 April 1982.

Water depth	Water temp. ^a	Photosynthetically available radiation ^b		Chlorophyll <i>a</i> ^c		Phytoplankton cell volume ^e	Dissolved inorganic carbon ^a
				uncorrected	corrected ^d		
(m)	(°C)	($\mu\text{E m}^{-2} \text{s}^{-1}$)	(%)	(mg m^{-3})	(mg m^{-3})	($\text{mm}^3 \text{m}^{-3}$)	(g m^{-3})
0	5.9	1247	100	2.47	1.76	211	30.9
1	5.9	756	60.0	2.60	1.90	258	30.5
2	5.8	459	36.8	2.56	1.87	375	30.4
3	5.8	307	24.6	2.42	1.78	329	30.3
4	5.8	205	16.4	—	—	—	—
5	5.7	143	10.8	2.50	1.83	273	29.7
8	5.6	51	4.1	2.31	1.69	182	31.0
10	5.6	26	2.1	2.25	1.64	154	29.7
12	5.6	14	1.14	—	—	—	—
15	5.6	5.7	0.46	1.93	1.40	161	29.8
18	5.6	2.4	0.19	—	—	—	—
Σ_{15}				34.69	25.36	3310	
\bar{x}				2.31	1.69	221	

^aMeasured by the staff at Limnological Institute, Konstanz.^bRecorded by Jewson *et al.* (1983).^cEnumerated by U. Sommer.^dCorrected for phaeopigments.**Table II.** Size and number of glass bottles used for exposure experiments at each exposure depth in two locations. For experimental type, see text.

Experimental type	Experimenter	Size (ml)	Number			Experimental locations
			light bottles	dark bottles	0-time blanks	
1 (i)	Tilzer	120	2	1	—	lake and pond
1 (ii)	Gächter and Maré	120	2	1	1 ^a	lake and pond
1 (iii)	Rai	570	2	1	—	lake
1 (iv)	Collos ^b	2000	1	0	1 ^c	lake
1 (iv)	Sakamoto	570	2	1	1 ^d	lake and pond
1 (v)	Tschumi <i>et al.</i>	120	3	3	3	lake
1 (v)	Dokulil	120	2	5 ^e	4 ^d	pond

^aOnly for the lake sample.^bFor simulated in situ exposure, the same size and number of bottles were employed at each light level.^cFor the samples at 0 m and 10 m depth.^dFor one integrated sample in the pond, but for each depth sample in the lake.^eAll dark bottles were incubated at a middle depth.

were used for *in situ* exposures in both lake and pond. Table II shows the size and number of the glass bottles used for these experiments. Details of the vertical perspex tube enclosures are described below.

Eight types of exposure experiments were run in parallel at both locations to estimate photosynthetic rates at each depth, P_z , and areal integral photosynthetic rates ΣP_z . (The investigators responsible for each method are also indicated below.)

Experiment 1. *In situ* exposure in suspended glass bottles. Lake water from different depths was incubated *in situ*. In the pond, samples from an integrated water column were exposed at 0, 0.15, 0.3, 0.5 and 1.0 m.

(i) Tilzer: Incubation with added ^{14}C -bicarbonate, followed by filtration on $0.8\ \mu\text{m}$ membrane filters. P_z was calculated from ^{14}C uptake values, total activity of added ^{14}C , and DIC concentration of lake waters (Gächter *et al.*, 1984).

(ii) Gächter and Marés: Subsamples were withdrawn from the same bottles used in Experiment (i) after incubation. The unfiltered samples were acidified, bubbled with air to remove inorganic ^{14}C , and their radioactivities determined. The activity of dissolved organic carbon in filtrates from parallel samples was also measured (Gächter *et al.*, 1984).

(iii) Rai: Lake samples from 0, 2 and 15 m of the lake were incubated with ^{14}C -bicarbonate, then filtered through 106, 35 and $10\ \mu\text{m}$ Nytal screens, respectively. Duplicate portions from these filtrations as well as from an unfiltered sample were subsequently passed through $0.2\ \mu\text{m}$ membrane filters and the contribution of each size fraction on total P_z was estimated (Rai, 1982).

(iv) Collos and Sakamoto: The ^{13}C method (Slawyk *et al.*, 1977; 1979) was used to measure P_z .

Collos: The working isotope solution was a ^{13}C bicarbonate solution ($3\ \text{mg C l}^{-1}$) prepared from $\text{Ba}^{13}\text{CO}_3$ (90.1 atom %: CEA, Service des Molecules Marquées, France), sterilized and stored in a 250 ml Teflon bottle. Each 2 l lake sample was incubated *in situ* with 6 ml of the isotope solution, and then filtered on Whatman GF/C glassfiber filters (47 mm diameter) under reduced vacuum ($<100\ \text{mm Hg}$). The filters were dried at 60°C and stored in a desiccator. Isotope analysis was carried out with a MS-10 mass spectrometer (Ass. Elect. Ind.) using each half of the filters. Particulate carbon of the lake water was determined with a Perkin-Elmer 240 C elementary analyser on 100 ml samples which were passed through Whatman GF/C filters (24 mm diameter) and dried at 60°C .

Sakamoto: $\text{NaH}^{13}\text{CO}_3$ (91.8 atom %: BOC. Ltd., London) was dissolved in glass distilled water, dispensed into pyrex ampoules and sterilized in boiling water. The final inorganic carbon concentration ($3.3\ \text{mg C ml}^{-1}$) and isotope ratio of the working solution were later determined with an infrared CO_2 analyser and a mass spectrometer. The water samples incubated with ^{13}C isotope solution and the zero time blank samples, taken just after an addition of isotope solution, were filtered onto preignited Whatman GF/C filters (24 mm diameter). The filters were dried, and particulate carbon content and isotope ratio were simultaneously determined with a small quadrupole mass spectrometer (Anelva, TE-150) coupled with an automatic elementary analyser (Yanagimoto, MT-2) after Otsuki *et al.* (1983).

P_z values were calculated similarly by Collos and Sakamoto from the atom % excess ^{13}C of filter samples, the atom % excess ^{13}C of the inorganic carbon pool, DIC, and particulate carbon concentration of the samples (Slawyk *et al.*, 1979).

(v) Tschumi *et al.* and Dokulil: In both the lake and the pond, light and dark bottles were incubated *in situ* to measure changes in dissolved oxygen, DO. For the lake experiments, Tschumi *et al.* employed an improved oxygen technique with a precision of $2-5\ \mu\text{g O}_2\ \text{l}^{-1}$ (Tschumi *et al.*, 1978), and in the pond series,

Dokulil used the back-titration and amperometric endpoint detection technique (Talling, 1973).

Experiment 2. Simulated in situ exposure experiment. Collos: Pyrex glass-bottles (2 l) were wrapped in neutral density nickel screens to give 60, 38, 21, 10, 6 and 2% of I_0 and filled with the lake water from the corresponding depths. After addition of ^{13}C solution, the bottles were placed in a water-cooled incubator on the Institute pier close to the lake station. Subsequent processing and analysis were as in Collos, Experiment 1(iv).

Experiment 3. In situ exposure in vertical perspex tubes. (i) Bossard and Uehlinger: Transparent plexiglass tubes (2 m length, 19 mm diameter), which can be linked together to give any desired length, were used for exposures. A thin silicon hose (1.3 mm inner diameter) of the same length as the exposure tube was filled with $\text{NaH}^{14}\text{CO}_3$ solution ($9 \mu\text{Ci ml}^{-1}$), and inserted with its base at the bottom end of the exposure tube prior to the start of experiment. After the whole system was placed vertically in the water column, a device at the bottom-end closed the tube bottom and also broke a glass tip at the base of the silicon hose (Figure 1). Homogeneous spiking of ^{14}C solution throughout the enclosed water

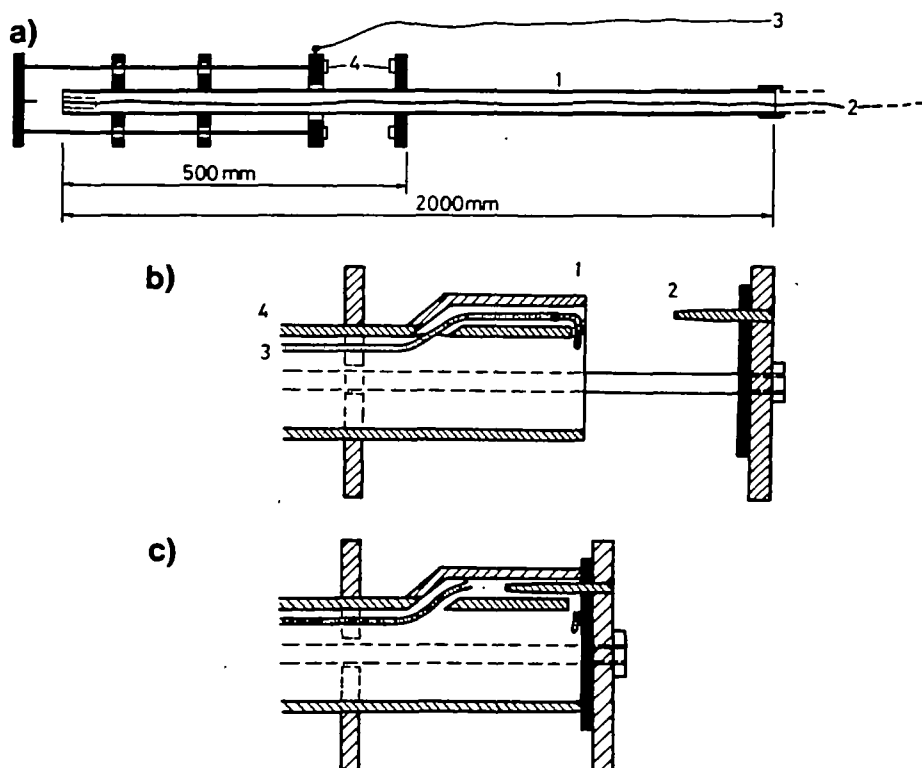


Fig. 1. Exposure tube (Bossard and Uehlinger) showing closure and release of ^{14}C (a) General view of exposure tube. (1) Tube; (2) Silicon hose; (3) Closing twine; (4) Magnets. (b) Tube as lowered into water in open position. (1) Tube wall; (2) Tube bottom with pin; (3) Silicon hose containing ^{14}C bicarbonate solution with glass tip. (c) Tube after closure and breakage of glass tip.

Table III. Exposure times and surface irradiance data during the exposure experiments in Lake Konstanz and algal pond on 21 April 1982. Light data are from Jewson *et al.* (1983).

Experimenter	Time	PAR (I_0) (Quanta cm^{-2})	Average PAR ($\mu\text{Einstein m}^{-2} \text{s}^{-1}$)
Lake			
Tilzer and Gächter	08.50–12.51	1.21×10^{21}	1386
Rai	08.55–12.55	1.15×10^{21}	1326
Sakamoto	09.15–13.15	1.16×10^{21}	1342
Collos	09.25–13.25	1.18×10^{21}	1362
Tschumi <i>et al.</i>	09.30–13.30	1.18×10^{21}	1360
Pond			
Tilzer, Gächter and Dokulil	10.32–14.32		
Sakamoto	10.40–15.00		

column was accomplished by pulling the thin hose upwards through a squeezing mechanism located at the upper end of the exposure tube. Final activity in the exposure tube was about $40 \mu\text{Ci l}^{-1}$. After light exposure, the contents of the tube were transferred into 5 l glass bottles, and the ^{14}C uptake was measured in subsamples by acid-bubbling and scintillation counting as in Experiment 1(ii).

(ii) Nusch: A vertical water column was enclosed by lowering transparent and darkened plexiglass tubes (1 or 2.5 m length, 40 mm inner diameter) into the water and incubated with or without vertical circulation by pump as described in Nusch (1981). Dissolved oxygen in the tube was measured with an oxygen probe (Chemtronic Waltemode OX653) installed in a measuring cell through which the incubated water passed. P_z and ΣP_z were estimated from differences in oxygen probe reading before and after exposure (Nusch, 1981), and, on some subsamples, by Winkler titration.

Mooring and exposure of the enclosed samples

At the lake station, five buoys were connected to each other at 3 m intervals with ropes, and bottles were suspended on lines from the center of each buoy so as to minimize shading effects. Four of the buoys were constructed of two floats connected by a long metal rod, one buoy was made of a triangle of plastic tubes. In the pond, sample bottles were hung from a long steel pole which crossed the pond, or were fixed at desired depths on a vertical stand. Tube samples were held vertically by special buoyant frames of transparent plexiglass (Nusch, 1981).

Exposure times and light conditions

The exposure times and I_0 during the experiments are given in Table III. Although there were some differences in light environment, these should not have been large enough to cause any significant changes in production rates between experiments.

Isotopic and chemical analysis of experimental samples and data work up

After the termination of exposure, all subsequent work up of samples (fil-

Comparisons of methods of production measurement

Table IV. Lake Konstanz: photosynthetic rates P_z , ($\text{mg C m}^{-3} \text{ h}^{-1}$, or $\text{mg O}_2 \text{ m}^{-3} \text{ h}^{-1}$) determined *in situ* by suspended bottle methods.

Depth (m)	P_z ($\text{mg C m}^{-3} \text{ h}^{-1}$)					P_z ($\text{mg O}_2 \text{ m}^{-3} \text{ h}^{-1}$)	
	Tilzer	Gächter	Rai	Collos	Sakamoto	Tschumi GP ^a	NP ^b
0	1.03	1.2	1.35	1.6	3.3	-2.2	-22.1
1	4.08	5.1	3.91	5.1	7.6	9.4	1.8
2	3.82	6.2	4.48	3.2	3.8	14.4	9.5
3	3.45	5.3	—	3.6	—	18.2	12.5
5	3.27	5.2	—	4.9	2.6	14.2	9.7
8	2.13	3.1	—	2.1	1.4	5.4	-1.1
10	1.12	1.3	—	—	—	0.1	-6.8
15	0.25	0.5	0.48	0.6	0.3	-7.6	-9.7
Method	¹⁴ C filtered samples	¹⁴ C acid- bubbling	¹⁴ C filtered samples	¹³ C mass- spectro- metry	¹³ C mass- spectro- metry	¹⁸ O ₂ precise photometric titration	

^aGP: Gross production

^bNP: Net production

Table V. Algal Pond: photosynthetic rates P_z , ($\text{mg C m}^{-3} \text{ h}^{-1}$, or $\text{mg O}_2 \text{ m}^{-3} \text{ h}^{-1}$) determined *in situ* by suspended bottle methods (Chl *a* content 12.5 mg m^{-3} , dissolved inorganic carbon 9.86 g C m^{-3}).

Depth (m)	PAR (%)	P_z ($\text{mg O}_2 \text{ m}^{-3} \text{ h}^{-1}$)			P_z ($\text{mg O}_2 \text{ m}^{-3} \text{ h}^{-1}$)	
		Tilzer	Gächter	Sakamoto	Dokulil GP	NP
0	100	2.4	4.1	4.5	150	140
0.15	87	3.5	4.7	4.9	70	60
0.30	77	3.5	5.3	5.2	60	50
0.50	68	4.7	6.5	5.6	50	40
1.00	53	5.3	7.7	6.1	50	40
Method		¹⁴ C filtered samples	¹⁴ C acid bubbling	¹³ C mass- spectro- metry	¹⁸ O ₂ amperometric end point detection	

tration, isotope and chemical analysis etc.) were done immediately at the Institute laboratory in order to permit initial evaluation by the GAP Workshop participants. (¹³C mass spectrometric analyses were an exception to this.)

Results and Discussion

The photosynthetic rates measured by suspended bottle methods in the lake and the experimental algal pond are shown in Tables IV and V. For replicated samples, the mean value at each depth was tabulated. P_z values from the ¹³C method were calculated by multiplying specific uptake rates, V_c , by particulate carbon concentrations (Table VI). Five vertical profiles of photosynthetic rates measured *in situ* at the lake were substantially similar (Figure 2), although differences were found in the depths of maximum activity, P_{max} , especially between

Table VI. Lake Konstanz: Specific carbon uptake rates (V_c)^a determined *in situ* by ^{13}C methods, and particulate carbon concentration.

Depth (m)	V_c (h^{-1})		Particulate carbon (g C m^{-3})	
	Collos	Sakamoto	Collos	Sakamoto
0	0.003	0.007	0.46	0.45
1	0.010	0.017	0.47	0.44
2	0.007	0.008	0.45	0.45
3	0.010	—	0.37	—
5	0.010	0.007	0.50	0.40
8	0.005	0.004	0.39	0.31
15	0.003	0.001	0.25	0.25

^aCalculated from the atom excess $\%^{13}\text{C}$ for the incubated samples, for the unenriched or zero time blank samples, and for ambient inorganic carbon in the lake water after Slawyk *et al.* (1979).

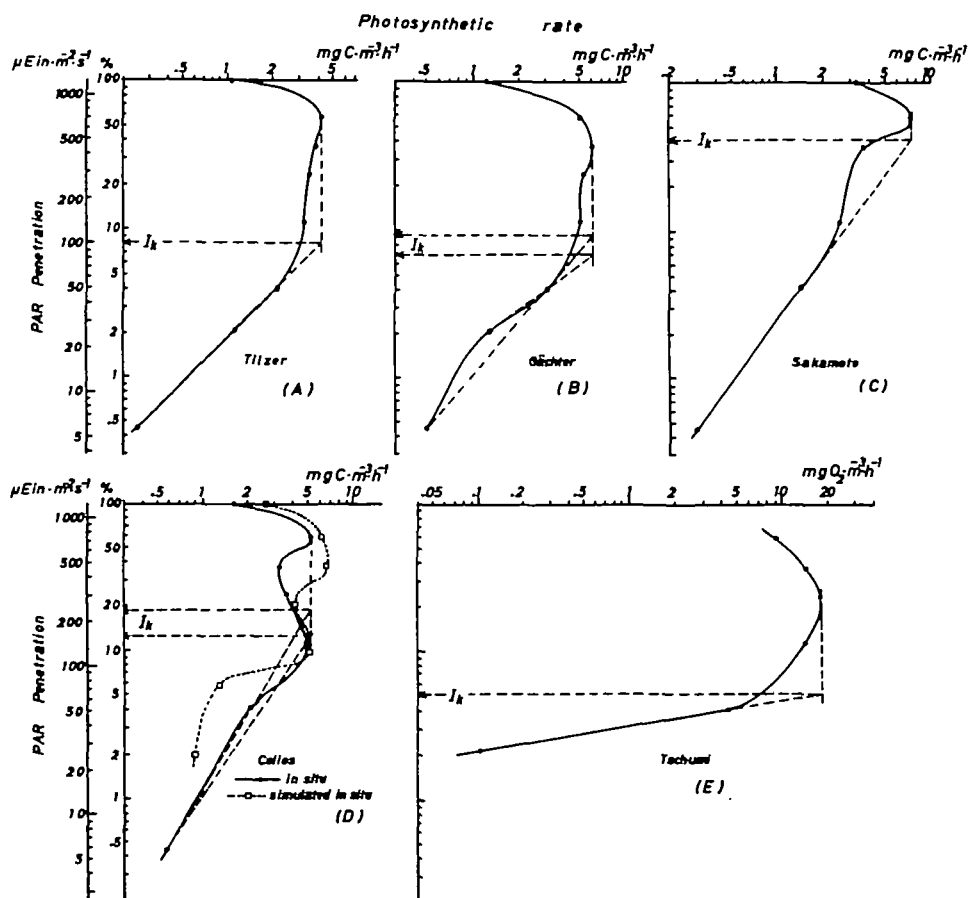


Fig. 2. Lake Konstanz: Profiles of primary productivity and estimated I_k measured by ^{14}C , ^{13}C and O_2 methods with *in situ* bottle incubation. (Note: the oxygen profile is based on gross photosynthetic measurements.) The sampling depths indicated are 0, 1, 2, 3, 5, 7, 10 and 15 m.

Comparisons of methods of production measurement

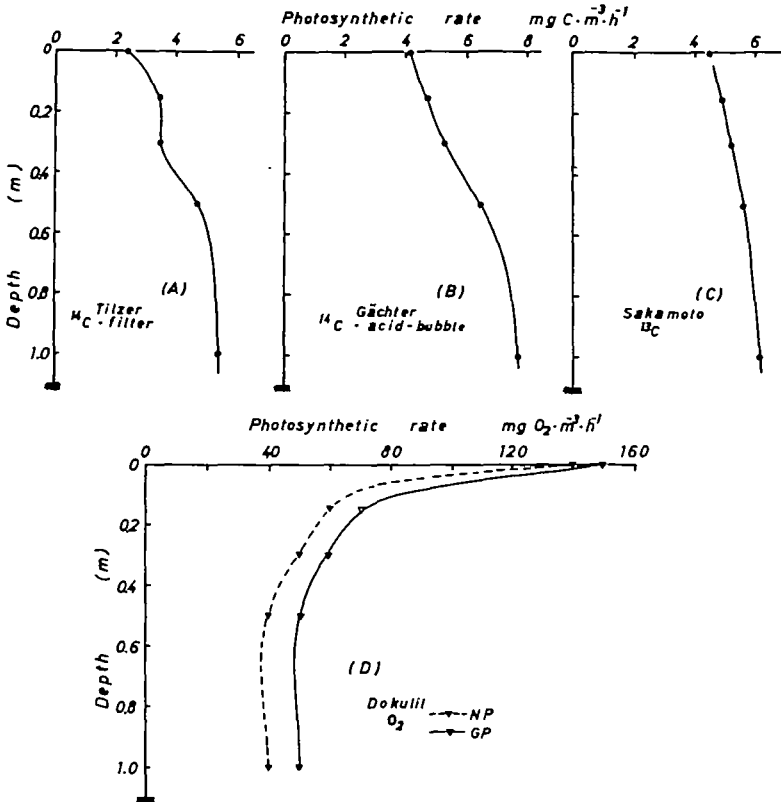


Fig. 3. Algal pond: profiles of primary production from ^{14}C , ^{13}C and oxygen methods.

the carbon uptake (^{14}C and ^{13}C) and DO methods.

P_z values given by the simulated *in situ* exposure with ^{13}C technique (Collos) converted to equivalent light depths are also shown in Figure 2. The main discrepancy between the *in situ* and simulated results was the depth of P_{max} .

At the pond, the profile derived from the oxygen bottle method was remarkably different from that given by ^{13}C and ^{14}C methods (Figure 3). All the isotope techniques gave P_{max} at the lowest depth in contrast to the surface maximum found with the oxygen method.

There was considerable variation for P_z values given by different ^{14}C and ^{13}C methods. In the lake, the greatest variation was noted at 0 and 1 m, (coefficient of variation c.v. = $\sigma/100/\bar{x}$), 70.8 and 65.4% respectively and lower c.v. (10.7–34.8%) for depths below 2 m. Similar variations were observed at the pond where the c.v. was 30.3% at the surface and 16.1–19.2% at the depths below 0.15 m.

Several reasons may be suggested for these variations. One cause could be the inconsistency of corrections for dark uptake and isotope discrimination used by different experimenters. In Experiment 1(i) and 1(iii), dark C uptake values were subtracted from light values at each depth, and also a correction for isotope dis-

crimination was applied, but in the other isotope experiments these corrections were not used. According to Gächter *et al.* (1983), dark uptake ranged from 8–17% of the light values in Experiments 1(i) and 1(ii) at the pond where the whole water column was under saturating insolation. In the ^{13}C experiments by Sakamoto in the lake, dark uptake was 1–2% of the light values near the surface and 5–9% at deeper layers: in the pond, dark uptake was 1.2–1.6%. Thus, inaccuracies in this correction factor might explain some discrepancies in the deeper P_z values but would have little effect on near surface estimates. The isotope discrimination factor proposed by Fisher *et al.* (1979), is smaller than the factor of 1.06 used here in Experiments 1(i), and 1(iii). No correction for isotope discrimination was applied for the present ^{13}C results but it is unlikely that this would explain the observed differences between ^{14}C and ^{13}C methods.

The variation between duplicate samples in any one single experiment was low. The relative deviation of P_z values to their mean was <12% on the results from Experiment 1(i), <16% on Experiment 1(ii) and <5% on Experiment 1(iii). Although the systematic errors in ^{14}C productivity work should not exceed 5% with careful blank control, standardization of ^{14}C working solution, and the correct evaluation of ambient DIC (Strickland, 1960), some differences of experimental manipulation may produce systematic errors with a c.v. up to 10% (Doty and Oguri, 1958). Even with a single technique, the c.v. may reach ~20% between measurements from successive samples (Doty *et al.*, 1965). This indicates that the variation found in the results between different experiments in this series was well within the systematic error to be expected in field work.

Maske (1984) examined the effect of ultraviolet radiation on photosynthetic carbon assimilation in these experiments and showed that ^{14}C uptake could increase because of protection from short wave length radiation when glass vessels were used as incubation enclosures. Similar results have been reported by Jitts *et al.* (1976) and Lorenzen (1979). The glass bottles used in our experiments were different in size (Table II) and came from various manufacturers. Some of the observed differences between experiments especially at higher irradiance, may have been due to such 'bottle effects'.

The range in experimental bottle size may also have led to initial differences in the enclosed plankton assemblages (e.g. in the numbers of zooplankton) and a subsequent heterogeneity of phytoplankton populations during incubation (Sheldon *et al.*, 1973; Venrick *et al.*, 1977).

Further causes of variations in P_z values between experiments can be suggested and should be carefully evaluated in future. In the ^{14}C methods, errors could arise in the standardization of ^{14}C working solution (Iverson *et al.*, 1976), because of heavy metal contamination (Carpenter and Lively, 1980), and in the quench correction factors used. Difference in the kinds of filters employed by each experimenter (membrane filters of 0.8 μm pore size in Experiment 1(i), of 0.2 μm in Experiments 1(ii) and 1(iii), and glass fiber filters of ~1.0 μm in Experiment 1(iv)) could have affected the relative amounts of retained particulate and filtrate ^{14}C . Gächter *et al.* (1984) detected a negligibly small release of labelled dissolved organic matter in 0.2 μm filtrates. 'Filtrate' losses would be expected to be higher from 0.8 μm and 1.0 μm filters leading to underestimations of the particulate

Table VII. Lake Konstanz: vertical changes of the Photosynthetic Quotient (PQ). Determined from oxygen production rates (Tschumi) and carbon uptake rates (filtered samples, Tilzer, and acid bubbling method, Gächter and Maréchal using ^{14}C technique).

Depth (m)	PQ	
	filtered	acid-bubbled
1	0.87	0.69
2	1.41	0.87
3	1.98	1.29
5	1.63	1.02
8	0.95	0.63
mean	1.37	0.90

production rates. In contrast, filter contamination by inorganic and dissolved organic matter would overestimate P_z (Arthur and Rigler, 1967; Berman and Eppley, 1974).

A source of error specific to the ^{13}C method is the possibility of incomplete combustion of the samples to CO_2 gas for mass spectrometric analysis which would lead to underestimations of the carbon uptake rate. Because of the systematic error involved in sample preparations and mass spectrometric analysis, the estimated P_z values at $\leq 1 \text{ mg C m}^{-3} \text{ h}^{-1}$ are of low reliability (e.g. 8 and 15 m depths).

There are inherent difficulties in comparing results of O_2 methods to those from ^{14}C and ^{13}C because the transformation of oxygen evolution to carbon uptake is not straightforward (McAllister *et al.*, 1964; Harris and Piccinin, 1977; Harris, 1980). However, if the values obtained by short term (up to 4–5 h) ^{14}C and ^{13}C methods are assumed to be close to gross photosynthesis (Dring and Jewson, 1982), then these may be compared to gross photosynthesis estimated by the O_2 method to evaluate the photosynthetic quotient, PQ (O_2 mol evolved/ CO_2 mol assimilated). An example for the lake station based on data from Experiments 1(i), 1(ii) and 1(v) (Tschumi) is given in Table VII. The estimated PQ values ranged from 0.80–1.98 and 0.63–1.29 for ^{14}C filtered and ^{14}C acid-bubbled methods, respectively. These results emphasize that the PQ values are not constant as has often been assumed, and as Williams *et al.* (1981) have stressed, will vary under different environmental and physiological conditions.

The net P_z as determined by the O_2 method should be somewhat underestimated due to the inclusion of respiration of zooplankton and heterotrophic microorganisms. However, if we assume that this effect was small during our experiments, the daytime compensation depth in the lake would be located at $\sim 8 \text{ m}$ (Table IV), corresponding to $\sim 4.0\%$ of I_0 but not at 12.5 m (1% I_0). The shallower compensation depth is compatible with the observations of Heaney and Sommer (1984) who followed the quantitative changes in phytoplankton biomass in the lake water samples (suspended *in situ*) during the joint field experiments and found only a very small increase in total cell volume at 7 m over 76 h.

Specific photosynthetic rates, P_z^B ($\text{mg C mg Chl a}^{-1} \text{ h}^{-1}$) in the lake were within the ranges often encountered in mesotrophic lakes (Table VIII) but P_z^B values in the pond were surprisingly low despite the eutrophic appearance of the

Table VIII. Lake Konstanz: specific photosynthetic rates P_z^B , (mg C mg Chl $a^{-1} h^{-1}$ or mg O₂ mg Chl $a^{-1} h^{-1}$) determined *in situ* by suspended bottle methods.

Depth (m)	(mg C mg Chl $a^{-1} h^{-1}$)					(mg O ₂ mg Chl $a^{-1} h^{-1}$)	
	Tilzer	Gächter	Rai	Collos	Sakamoto	GP	Tschumi NP
0	0.58	0.68	0.77	0.91	1.87	-1.22	-12.56
1	2.17	2.68	2.06	2.68	3.68	4.93	0.95
2	2.04	3.32	2.40	1.71	2.03	7.67	5.09
3	1.94	2.98	—	2.02	—	10.20	7.02
5	1.78	2.84	—	2.68	1.42	7.75	5.27
8	1.24	1.83	—	1.24	0.83	3.18	-0.66
10	0.68	0.79	—	—	—	0.01	-4.16
15	0.18	0.36	0.34	0.43	0.21	-5.39	-6.89

Table IX. Algal Pond: specific photosynthetic rates P_z^B , (mg C mg Chl $a^{-1} h^{-1}$ or mg O₂ mg Chl $a^{-1} h^{-1}$) determined *in situ* by suspended bottle methods.

Depth (m)	(mg C mg Chl $a^{-1} h^{-1}$)			(mg O ₂ mg Chl $a^{-1} h^{-1}$)	
	Tilzer	Gächter	Sakamoto	GP	Dokulil NP
0	0.19	0.33	0.36	12.0	11.2
0.15	0.28	0.38	0.39	5.6	4.8
0.30	0.28	0.42	0.42	4.8	4.0
0.50	0.38	0.52	0.45	4.0	3.2
1.00	0.32	0.62	0.49	4.0	3.2

green water (Table IX). The pond samples had a low concentration of nutrients ($PO_4\text{-P} < 5 \mu\text{g l}^{-1}$, total-N 0.3 mg l^{-1}), and the phytoplankton population which had peaked a few days prior to the experiment was probably in a senescent stage.

Areal production rates (obtained by integration of P_z down to indicated depths) are given in Tables X and XI for the lake and pond, respectively. In the lake ΣP_z values derived from ^{14}C filtration and ^{13}C methods were fairly close for the water column deeper than 5 m. The acid-bubbling ^{14}C method (Gächter) gave somewhat higher ΣP_z values for the deeper part of the water column. The shallower the water column, the greater were the differences in ΣP_z estimates. Areal rates calculated from O₂ experiments were compatible with those obtained with ^{14}C as described later and indicated that, in the lake, net production was <50% of gross production in the euphotic zone.

It was of interest to compare ΣP_z estimated from the suspended bottle methods with those obtained from the tube exposures. The latter techniques are of specific advantage for ΣP_z because they give a complete integration of the entire water column, and also because they are relatively simple to handle. One disadvantage, however, is that at present such methods are suitable only for comparatively shallow water columns.

The P_z values determined in vertical tube exposure followed by the acid-bubbling ^{14}C technique (Experiment 3(i)) and those based on the measurement by

Table X. Lake Konstanz: areal photosynthetic rates ΣP_z , ($\text{mg C m}^{-2} \text{h}^{-1}$, or $\text{mg O}_2 \text{m}^{-2} \text{h}^{-1}$) determined *in situ* by suspended bottle methods.

Integration depth (m)	$(\text{mg C m}^{-2} \text{h}^{-1})$					$(\text{mg O}_2 \text{m}^{-2} \text{h}^{-1})$	
	Tilzer	Gächter	Rai	Collos	Sakamoto	Tschumi GP	NP
down to							
1	2.6	3.2	2.63	3.4	5.5	3.6	-10.2
2	6.5	8.8	6.83	7.5	11.2	15.5	-4.5
3	10.3	11.4	—	10.9	(14.4)	31.7	6.6
5	17.0	23.1	—	19.4	20.2	64.1	28.7
8	25.1	35.5	—	29.9	26.2	93.4	41.5
10	28.3	39.9	—	(32.6)	(27.9)	98.9	33.5
15	31.7	45.4	—	39.4	32.1	80.3	-7.7

Table XI. Algal Pond: areal photosynthetic rates ΣP_z , ($\text{mg C m}^{-2} \text{h}^{-1}$, or $\text{mg O}_2 \text{m}^{-2} \text{h}^{-1}$) determined *in situ* by suspended bottle methods.

Integration depth (m)	$(\text{mg C m}^{-2} \text{h}^{-1})$			$(\text{mg O}_2 \text{m}^{-2} \text{h}^{-1})$	
	Tilzer	Gächter	Sakamoto	GP	Dokulil NP
down to					
0.15	0.44	0.64	0.71	16.5	15.0
0.25	(0.79)	(1.30)	(1.11)	(23.0)	(25.0)
0.30	0.97	1.57	1.36	26.2	30.0
0.50	1.79	2.53	2.44	37.2	39.0
0.75	3.04	4.24	3.91	49.7	49.0
1.00	4.26	6.08	5.37	62.2	59.0
1.10	(4.79)	(6.85)	(5.98)	(67.7)	(63.0)

oxygen electrode or Winkler titration (Experiment 3(ii)) are given in Tables XII and XIII respectively. Compared with the carbon isotope techniques (Tables X and XI), the ^{14}C tube method consistently gave higher ΣP_z values. Harris and Piccinin (1977) showed that high midday irradiation can depress net photosynthesis measured in bottles *in situ* by 20 to 80%. The higher ΣP_z values observed in tube exposures in the lake may be partly due to the ability of the algae to migrate throughout the enclosed water column (Marra, 1978) and may give a better measurement of ΣP_z than those from suspended bottle experiments.

In the pond, the ΣP_z value obtained from the profiled exposure tube was in close agreement with that from the suspended bottle experiment despite some difference of exposure times (Table XIII). Also, the Winkler titration on water circulated in the tubes gave similar estimates to those by DO electrode. However, in the lake, where the changes of DO content during the exposure period were small, ΣP_z estimates by DO electrode in the tube exposure were almost four times those measured with the precise suspended oxygen bottle method. This suggests that at present the tube exposure with DO electrode as a sensor is limited in use mainly to high productivity waters.

All the data on the areal production rates in the euphotic zone (ΣP_{eu}) are summarized in Table XIV. In the lake, ΣP_{eu} is given (1) for the water column down to

Table XII. Photosynthetic carbon fixation rates determined in vertical tube exposures using the ^{14}C method (Bossard and Uehlinger). Exposure time was 11:28–15:40 (4.2 h) at the lake and 12:03–17:03 (5 h) at the pond.

Location	Depth (m)	P_z ($\text{mg C m}^{-2} \text{h}^{-1}$)	water column (m)	ΣP_z ($\text{mg m}^{-2} \text{h}^{-1}$)
Lake	0–1	5.7	down to 1	5.7
	1–2	6.4	2	12.1
	2–3	6.5	3	18.6
	3–4	9.1	4	27.7
Pond	0–0.25	6.4	down to 0.25	1.6
	0.25–0.50	6.7	0.50	3.3
	0.50–0.75	8.0	0.75	5.3
	0.75–1.10	10.7	1.10	8.3

Table XIII. Comparison of areal gross production rates ($\text{mg O}_2 \text{m}^{-2} \text{h}^{-1}$) estimated from the changes of DO content in vertical perspex tubes and in suspended glass bottles in Lake Konstanz and the algal pond on 21 April 1982. Exposure times for the tube and suspended bottle experiments were 9.30–17.00 h and 10.32–14.32 h in the pond, and 11.15–45 h and 9.30–13.30 h in the lake, respectively.

Measured with	Exposure type	$(\text{mg O}_2 \text{m}^{-2} \text{h}^{-1})$	
		Lake 2 m ^a	Pond 1 m ^a
DO electrode ^b	Vertical tube with discrete profiling	61	65
	Vertical tube with circulation	—	32
Winkler titration	Vertical tube with discrete profiling	—	54 ^c
	Vertical tube with circulation	—	36 ^c
	Suspended bottles	16 ^d	62 ^e

^aDepth of water column enclosed in tubes.^bNusch.^cNusch and Dokulil.^dTsuchumi *et al.*^eDokulil.

8 m i.e., the daytime compensation depth estimated from O_2 profiles and (2) down to 15 m, close to the limit of gross production. If we use $\text{PQ} = 1$ (but see reservations above) the suspended bottle oxygen method gave a gross ΣP_{eu} of 35.0 and 30.1 $\text{mg C m}^{-2} \text{h}^{-1}$ for 8 and 15 m water columns respectively. At the pond, the calculated values of gross P_{eu} down to 1.0 m depth (23.3 $\text{mg C m}^{-2} \text{h}^{-1}$ for suspended bottles and 24.4 $\text{mg C m}^{-2} \text{h}^{-1}$ for tube exposures) were considerably higher than those derived from the bottle exposures with ^{14}C and ^{13}C methods (Table XI) or from the ^{14}C -tube exposure. Thus, the correlation between oxygen and carbon methods in suspended bottles was reasonable for the lake experiments but poor for the pond. In part, these differences could have been due to the senescent state of the algae in the pond which could cause a large error in our assumed PQ value (see above).

An important algal parameter is I_k , the light intensity at which an extension from the initial slope of the P versus I curve intersects with the P_{max} (Talling, 1957). The I_k values were estimated graphically for five profiles of P_z versus downwelling light penetration (Figure 2). The lowest estimate of I_k was 65 μEins .

Table XIV. Summarized areal production rates for euphotic zone in Lake Konstanz and Algal Pond.

Experimenter	Method and enclosure	Lake			Pond		
		ΣP_{Lm} (mg C m ⁻² h ⁻¹)	ΣP_{Lm} (mg C mg Chl a ⁻¹ h ⁻¹)	ΣP_{Lm} (mg C m ⁻² h ⁻¹)	ΣP_{Lm} (mg C mg Chl a ⁻¹ h ⁻¹)	ΣP_{Lm} (mg C m ⁻² h ⁻¹)	ΣP_{Lm} (mg C mg Chl a ⁻¹ h ⁻¹)
Tilzer	¹⁴ C, filtered (120 ml)	25.1	1.73	31.7	1.25	4.79	0.35
Gächter	¹⁴ C, acid-bubbling (120 ml)	35.5	2.24	45.4	1.79	6.85	0.50
Collos	¹⁴ C, (2 l)	29.9	2.06	39.4	1.55	—	—
Sakamoto	¹⁴ C, (lake 0.5 l, pond 120 ml)	26.2	1.81	32.1	1.26	5.98	0.43
Bossard and Uehlinger	¹⁴ C, tube and acid-bubbling	—	—	—	—	8.30	0.60
<hr/>							
Tschumi	O ₂ improved GP precise technique GP (120 ml)	93.4	6.44	98.9 ^a	5.57 ^a	—	—
Dokull	O ₂ NP	41.5	2.86	33.5 ^a	1.89 ^a	—	—
	ampero-metric GP	—	—	-7.7	-0.30	—	—
	back-titration (120 ml)	—	—	—	—	62.2 ^b	4.98 ^b
	tube oxygen electrode	—	—	—	—	59.0 ^b	4.72 ^b
Nusch	O ₂ Winkler titration	—	—	—	—	65.0 ^b	5.20 ^b
		—	—	—	—	54.0 ^b	4.32 ^b

^aFor the column down to 10 m.

^bFor the column down to 1 m.

Table XV. Lake Konstanz: contribution of different size fractions in photosynthetic carbon assimilation, as measured with ^{14}C method by H. Rai.

Depth (m)	Size fractions (μm)	Carbon assimilation ($\text{mg C m}^{-2} \text{ h}^{-1}$)	%
0	<250 - >106	0.334	24.8
	<106 - > 35	0.094	7.0
	< 35 - > 10	0.209	15.5
	< 10	0.710	52.7
1	<250 - >106	0.677	17.3
	<106 - > 35	0.371	9.5
	< 35 - > 10	0.546	14.0
	< 10	2.314	59.2
2	<250 - >106	0.639	14.2
	<106 - > 35	0.256	5.7
	< 35 - > 10	0.643	1.0
	<	3.546	79.1

tein $\text{m}^{-2} \text{ s}^{-1}$ (5.2% I_0) from the improved oxygen method (Experiment 1(v), Tschumi). The ^{14}C methods gave fairly similar I_k , 102 $\mu\text{Einstein m}^{-2} \text{ s}^{-1}$ (Experiment 1(i)), 87 $\mu\text{Einstein m}^{-2} \text{ s}^{-1}$ or 147 (Experiment 1(ii)). Higher I_k s were obtained from the ^{13}C experiments, 158 or 237 $\mu\text{Einstein m}^{-2} \text{ s}^{-1}$ (Collos) and 524 - $\mu\text{Einstein m}^{-2} \text{ s}^{-1}$ (Sakamoto).

Some of the above variability is certainly due to the difficulty of determining I_k graphically with the available data. Also, there were differences in the handling of the experimental samples, especially in the care which was taken in shielding them from stray or excess light. Attention to methodological details must be emphasized if accurate assessments of parameters such as I_k and compensation depth are to be made.

Important supplementary information was added by the experiment in which ^{14}C photosynthetic fixation by different size fractions of the phytoplankton was measured (Rai, 1982). As has been observed for many other aquatic environments (e.g. Malone, 1980; Kalff, 1972; Pollinger and Berman, 1982), the nanoplankton (<10 μm) were the dominant producers in Lake Konstanz at this season (Table XV). Nanoplankton have high metabolic activities and are also the major food organisms for herbivorous zooplankton. For a comprehensive analysis of phytoplankton population dynamics and of ecosystem metabolism, the size partitioning of carbon assimilation is of particular importance. However, to be fully meaningful, this data should be accompanied by a parallel analysis of the biomass composition and growth rate at the species level (see Heaney and Sommer, 1984).

Conclusions

The GAP Workshop gave us an opportunity to compare several methods for measuring primary productivity in a simultaneous fashion. As we have shown, considerable variation between methods was found for P_z which sometime gave very different depth profiles (e.g. Figure 3). Despite this, the estimates by different methods of ΣP_z were reasonably similar. Techniques which use tube en-

closures appear to be promising but are limited to shallow water columns. The good correspondence found between ^{14}C and ^{13}C methods was very encouraging especially as the latter technique is less of a health hazard and can be particularly useful in aquatic environments where radioactive isotopes are legally prohibited.

Oxygen methods, especially new modifications which permit high sensitivity and precision, will undoubtedly continue to be widely used. Our experimental series emphasize the difficulties of relating O_2 measurements to those based on carbon isotope uptake, particularly with respect to evaluating PQ.

To our knowledge this is the first time that such a wide ranging comparison of field techniques for measuring algal photosynthesis has been made. Although the experiments were limited to two locations on a single day, we suggest that the results obtained are important for an assessment of the reliability (and weakness) of current methodologies for determining aquatic primary production.

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